

Proceedings of the Estonian Academy of Sciences 2025, **74**, 1, 15–22

https://doi.org/10.3176/proc.2025.1.02

www.eap.ee/proceedings Estonian Academy Publishers

ENVIRONMENTAL CHEMISTRY

RESEARCH ARTICLE

Received 27 October 2023 Accepted 18 July 2024 Available online 15 January 2025

Keywords:

proton transfer model, equilibrium system of $H_2O-CO_2-CaCO_3-NaH_2PO_4$, carbonate equilibrium, phosphate equilibrium

Corresponding author:

Ivar Zekker ivar.zekker@ut.ee

Citation:

Zekker, I., Artemchuk, O., Rikmann, E., Uiga, K., Daija, L., Mamun, F.-A. et al. 2025. Modelling of the closed equilibrium system $H_2O-(CO_2)_W-CaCO_3-NaH_2PO_4$ for wastewater treatment applications. *Proceedings of the Estonian Academy of Sciences*, **74**(1), 15–22. https://doi.org/10.3176/proc.2025.1.02

© 2025 Authors. This is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (http://creativecommons.org/licenses/by/4.0).

Modelling of the closed equilibrium system $H_2O-(CO_2)_W$ -CaCO₃-NaH₂PO₄ for wastewater treatment applications

Ivar Zekker^a, Oleg Artemchuk^a, Ergo Rikmann^a, Kalev Uiga^a, Laura Daija^a, Faysal-Al Mamun^a, Andrejs E. Krauklis^{b,c}, Ruta Ozola-Davidane^c, Taavo Tenno^a and Toomas Tenno^{†a}

- ^a Institute of Chemistry, University of Tartu, Ravila 14a, 50411 Tartu, Estonia
- ^b Laboratory of Advanced and Sustainable Engineering Materials, Department of Manufacturing and Civil Engineering (ASEMIab), Norwegian University of Science and Technology (NTNU), Building B, Teknologiveien 22, 2815 Gjøvik, Norway
- c Faculty of Geography and Earth Sciences, Department of Environmental Protection, University of Latvia, Jelgavas 1, LV-1004 Riga, Latvia

ABSTRACT

Water pollution caused by nitrogen and phosphorus compounds triggers algal blooms and poses challenges to aquatic organisms. Wastewater treatment applications represent critical areas where fundamental models of phosphorus equilibria, prevalent in municipal and reject water streams, are essential. This paper focuses on studying the closed equilibrium system H₂O- $(CO_2)_W$ -CaCO₃-NaH₂PO₄, elucidating the structural distribution of ions and molecules within this system. Utilizing the iteration method, we calculate pH, concentrations of formed ions and molecules, and proton transfer parameters based on a developed proton transfer model. Upon the initial formation of the equilibrium system H₂O- $(CO_2)_W$ -CaCO₃, CO_3^{-2} ions released from the dissolution of CaCO₃ bind protons (Δ [H⁺]_{CO₃}) originating from the reversible dissociation of water (Δ [H⁺]_{H₂O₃) and H₂CO₃ (Δ [H⁺]_{H₂CO₃). In the equilibrium system H₂O- $(CO_2)_W$ -CaCO₃-NaH₂PO₄, at low initial CO₂ concentrations (<3.39 × 10⁻⁵ mmol·L⁻¹), the majority of protons (Δ [H⁺]_{H₂PO₄) originate from the H₂PO₄ dissociation. Conversely, with increased initial CO₂ concentrations, the dissociation of H₂CO₃ (Δ [H⁺]_{H₂CO₃) becomes the primary proton source. Experimental validation confirms the developed model's accuracy.}}}}

1. Introduction

The decline in water quality and pollution resulting from nitrogen and phosphorus species are linked to algal blooms, leading to the damage and subsequent death of aquatic organisms and health issues for humans. Wastewater treatment systems require fundamental approaches to remove phosphorus species from municipal and wastewater streams, necessitating optimization of biological and chemical treatment systems, which, in turn, requires mathematical modelling of the related processes. The system $H_2O-(CO_2)_W$ -CaCO₃ served as the foundation for modelling the closed system H₂O-(CO₂)_W-CaCO₃-NaH₂PO₄ [1]. Models for the equilibrium system of $H_2O_{-}(CO_2)_{W}$ -CaCO₃, based on proton-centred acid-base equilibria, have been developed and published for both open [2] and closed systems [1–4]. Similar models have also been developed for the CaS-H₂O and SrS-H₂O systems [5–8]. These models can primarily be utilized in environmental engineering and for assessing greenhouse gas emissions [2]. The current model can be employed in wastewater treatment to estimate bacterial inhibition based on different substrate concentrations. In an aqueous medium, all acid-base equilibria are interconnected through proton concentration, meaning that a change in the concentration of any component of the system affects the concentrations of all components. In the equilibrium system H₂O- $(CO_2)_W$ -CaCO₃, reducing the concentration of $[CO_2]_W$ within the range of 3.39×10^{-5} mmol·L⁻¹ – 3.39×10^{-7} mmol·L⁻¹ leads to an increase in pH up to 9.88 [1,3]. Consequently, protons bind to CO_3^{2-} ions, resulting in an increase

in the concentration of $[OH^-]$ ions in the medium. At lower concentrations of $[CO_2]_W$, water dissociation provides the majority of protons for the equilibrium system, while the dissociation of H_2CO_3 can be neglected at small concentrations of $[CO_2]_W$.

Equilibrium systems comprising carbonaceous and phosphorus species are prevalent in various wastewater sources [9,10]. In instances such as landfill leachate and reject water from anaerobic digestion of organic substrates, wastewater often contains elevated concentrations of phosphorus ions $(H_2PO_4^- \text{ and } HPO_4^{2-})$. For example, total phosphorus concentrations can reach up to 3.22 mmol·L⁻¹ in landfill leachate [11–13] and 1.61–3.22 mmol·L⁻¹ in reject water [14,15], originating from the dewatering of digested sludge or manure from anaerobic methane digesters [15,16].

The present study aims to develop a model for the closed equilibrium system $H_2O-(CO_2)_W-CaCO_3-NaH_2PO_4$, utilizing iteration and square root equations, to ascertain equilibrium pH, concentrations of all formed ions and molecules, and proton transfer parameters at equilibrium state. The model's accuracy was further validated through experimental dissolution and potentiometric methods.

2. Materials and methods

2.1. Modelling the system H₂O-(CO₂)_W-CaCO₃-NaH₂PO₄

When modelling the self-regulating complex system $H_2O-(CO_2)_W-CaCO_3-NaH_2PO_4$, it is crucial to consider that acid-base reactions in aqueous solutions proceed rapidly. For instance, the dissociation reaction of water into H⁺ and OH⁻ ions is recognized as one of the fastest reaction rates in aqueous solutions, with a rate constant of $k = 1.4 \times 10^{11} \text{ L} \cdot \text{mol}^{-1} \cdot \text{s}^{-1}$ [17,18]. The calculations' results prior to the addition of NaH₂PO₄ to the H₂O-(CO₂)_W-CaCO₃ system are presented in Table 1 [1]. The term $\Delta[\text{H}^+]_{\text{H}_2\text{CO}_3}$ denotes the concentration of protons released from the dissociation of carbonic acid, and the term $\Delta[\text{H}^+]_{CO_3^{-2}}$ represents the concentration of protons bonded to carbonate ions released by dissolving CaCO₃.

The data presented in Table 1 illustrate that within the closed equilibrium system of $H_2O_{(CO_2)_W}$ –CaCO₃, the pH value reaches a maximum of 9.88 as the concentration of $[CO_2]_{W0}$ decreases [1]. At lower concentrations of $[CO_2]_{W0}$, the dissociation of water primarily contributes to proton formation, while the dissociation of H_2CO_3 can be disregarded due to the minimal proton production. As observed from Table 1, when $[CO_2]_{W0}$ is below 3.39×10^{-4} mmol·L⁻¹, the concentration of $[HCO_3^-]$ ions matches that of $[H^+]$ ions generated by water dissociation. Consequently, the equilibrium of the system is determined by the dissociation of water and the solubility of CaCO₃.

The distribution of ions and molecules within the $H_2O-(CO_2)_W-CaCO_3-NaH_2PO_4$ system aligns quantitatively with the equilibrium constants depicted in the structural scheme (Fig. 1). The designations of the symbols in Fig. 1 are as follows:

- $K_{\rm h}$ is the hydration constant of CO₂
- *K*'_{a1} is the dissociation constant of H₂CO₃
- K_{a2} is the dissociation constant of HCO₃⁻
- $K_{\rm SP}$ is the solubility product constant of CaCO₃
- $K_{\rm W}$ is the ion-product constant of water
- $K_{a2,P}$ is the dissociation constant of $H_2PO_4^-$
- $K_{a3,P}$ is the dissociation constant of HPO₄²⁻

Table 1. Results of the calculated proton mass-transfer and pH values, along with the concentrations of ions and molecules after $CaCO_3$ was added into the system $H_2O-(CO_2)_{WO}$ at T = 25 °C

[CO ₂] _{W0}	$\Delta [\mathrm{H}^+]_{\mathrm{H}_{2}\mathrm{O}}$	$\Delta [\mathrm{H}^+]_{\mathrm{H}_2\mathrm{CO}_3}$	$\Delta[\mathrm{H^{+}]_{\mathrm{CO}_{3}^{2-}}}$	Equilibrium pH	[Ca ²⁺]	[HCO ₃ ⁻]
mmol· L^{-1}	mmol·L ⁻¹	mmol·L ⁻¹	$mmol \cdot L^{-1}$		mmol·L ⁻¹	mmol·L ⁻¹
3.39×10^{-7}	7.63×10^{-2}	-2.03×10^{-5}	-7.63×10^{-2}	9.88	1.03×10^{-1}	7.63×10^{-2}
3.39×10^{-5}	7.62×10^{-2}	1.09×10^{-4}	-7.63×10^{-2}	9.88	1.03×10^{-1}	7.63×10^{-2}
3.39×10^{-4}	7.56×10^{-2}	7.00×10^{-4}	-7.63×10^{-2}	9.87	1.03×10^{-1}	$7.70 imes 10^{-2}$
1.32×10^{-2}	6.19×10^{-2}	1.57×10^{-2}	-7.75×10^{-2}	9.79	1.04×10^{-1}	9.32×10^{-2}
3.39×10^{-1}	2.43×10^{-3}	3.46×10^{-1}	-3.49×10^{-1}	8.38	3.56×10^{-1}	6.95×10^{-1}
2.45	-1.55×10^{-7}	1.73	-1.73	7.00	1.73	3.46
33.9	-1.11×10^{-3}	5.78	-5.78	5.95	5.78	11.6



Fig. 1. The structural scheme of the equilibrium distribution of ions and molecules in the final closed equilibrium system $H_2O-[CO_2]_W-CaCO_3-NaH_2PO_4$.

The equilibrium constants of the system under investigation are represented by Eqs (1-7):

$$K_{\rm h} = \frac{[{\rm H}_2 {\rm CO}_3]}{[{\rm CO}_2]_{\rm W}} = 2.58 \times 10^{-3} \, [19] \,, \tag{1}$$

$$K'_{a1} = \frac{[H^+] \times [HCO_3^-]}{[H_2CO_3]} = 1.72 \times 10^{-4} \text{ mol} \cdot L^{-1} [20], \qquad (2)$$

$$K_{a2} = \frac{[\mathrm{H}^+] \times [\mathrm{CO}_3^{2-}]}{[\mathrm{HCO}_3^{-}]} = 4.7 \times 10^{-11} \text{ mol} \cdot \mathrm{L}^{-1} [21], \qquad (3)$$

$$K_{\rm W} = [{\rm H}^+] \times [{\rm OH}^-] = 1.01 \times 10^{-14} \ ({\rm mol} \cdot {\rm L}^{-1})^2 \ [22], \tag{4}$$

$$K_{\rm SP} = [{\rm Ca}^{2+}] \times [{\rm CO}_3^{2-}] \cong 2.8 \times 10^{-9} \,\,({\rm mol} \cdot {\rm L}^{-1})^2 \,[23],\tag{5}$$

$$K_{a2,P} = \frac{[\text{HPO}_4^{2-}] \times [\text{H}^+]}{[\text{H}_2 \text{PO}_4^-]} \cong 6.2 \times 10^{-8} \text{ mol} \cdot \text{L}^{-1} [23],$$
(6)

$$K_{a3,P} = \frac{[PO_4^{3-}] \times [H^+]}{[HPO_4^{2-}]} \cong 4.2 \times 10^{-13} \text{ mol} \cdot L^{-1} \text{ [24]}.$$
 (7)

Taking into account that in the equilibrium system $H_2O-(CO_2)_W-CaCO_3-NaH_2PO_4$ the concentration of $[HPO_4^{2-}-P]$ exceeds the concentration of $[PO_4^{3-}-P]$ by a factor of several orders of magnitude ($[HPO_4^{2-}]>>[PO_4^{3-}]$), the charge balance equation can be expressed as:

$$2[Ca^{2+}] + [H^+] + [Na^+] \cong 2[CO_3^{2-}] + [HCO_3^-] + [OH^-] + [H_2PO_4^-] + 2[HPO_4^{2-}].$$
(8)

The molar balance equation of carbonaceous species can be expressed as follows:

$$\Sigma[CO_X]_W = \Sigma[CO_X]_{W0} + [Ca^{2+}] = [CO_3^{2-}] + [HCO_3^{-}] + [H_2CO_3] + [CO_2]_W, \qquad (9)$$

where $\sum [CO_X]_W$ is the total concentration of all carbonaceous species. It is noteworthy that this equation does not change when NaH₂PO₄ is supplemented.

Taking into account that in the equilibrium state of the system the concentration of $[HPO_4^{2-}-P]$ exceeds the concentration of $[PO_4^{3-}-P]$ by a factor of several orders of magnitude ($[HPO_4^{2-}] >> [PO_4^{3-}]$), the molar balance equation of phosphorus-containing particles can be expressed as follows:

$$c[NaH_2PO_4] \cong [H_2PO_4^-] + [HPO_4^{2^-}] \cong [Na^+] = const,$$
 (10)

18 I. Zekker et al.

where $c[NaH_2PO_4]$ represents the concentration of supplemented NaH_2PO_4 to the equilibrium system $H_2O_{-}(CO_2)_W$ -CaCO₃.

To calculate the concentrations of $[CO_2]_W$, $[CO_3^{2-}]$, $[H_2CO_3]$, $[HCO_3^{-}]$, $[H^+]$, $[OH^-]$, $[Ca^{2+}]$, $[H_2PO_4^{-}]$ and $[H_2PO_4^{2-}]$ in the final equilibrium system, it is necessary to solve nine equations (1–9). To reduce the number of equations, the composite constant K_{a1} is utilized. The first pseudo-acid ionization constant K_{a1} [11] is expressed as the product of the constants K_h and K_{a1} :

$$K_{a1} = K_{h} \times K'_{a1} = 4.44 \times 10^{-7} \,\text{mol} \cdot \text{L}^{-1}.$$
(11)

Based on Eqs (6) and (10), the following dependencies were developed:

$$[H_2PO_4^-] = [Na^+] \times \frac{[H^+]}{[H^+] + K_{a2,P}}$$
(12)

and

$$[HPO_4^{2-}] = [Na^+] \times \frac{K_{a2,P}}{[H^+] + K_{a2,P}}.$$
(13)

Replacing the variables from Eqs (3), (4), (5), (11), (12), and (13), Eqs (8) and (9) can be expressed as follows:

$$2 \times \frac{K_{\rm SP}}{K_{\rm a2}} \times \frac{[{\rm H}^+]}{[{\rm HCO}_3^-]} + [{\rm H}^+] = 2 \times K_{\rm a2} \times \frac{[{\rm HCO}_3^-]}{[{\rm H}^+]} + [{\rm HCO}_3^-] + \frac{K_{\rm W}}{[{\rm H}^+]} + [{\rm Na}^+] \times \frac{K_{\rm a2,P}}{[{\rm H}^+] + K_{\rm a2,P}},$$
(14)

$$\Sigma[CO_{X}]_{W0} + \frac{K_{SP}}{K_{a2}} \times \frac{[H^{+}]}{[HCO_{3}^{-}]} = K_{a2} \times \frac{[HCO_{3}^{-}]}{[H^{+}]} + [HCO_{3}^{-}] + [HCO_{3}^{-}] \times \frac{[H^{+}]}{K_{a1}}.$$
 (15)

As a result of replacing variables, the system is characterized by two equations with two unknown concentrations $[HCO_3^-]$ and $[H^+]$. Equations (14) and (15) are converted to the following forms:

$$\left(\frac{2 \times K_{a2}}{[H^+]} + 1\right) \times [HCO_3^-]^2 + \left(\frac{K_W}{[H^+]} - [H^+] + [Na^+] \times \frac{K_{a2,P}}{[H^+] + K_{a2,P}}\right) \times [HCO_3^-] - \frac{2 \times K_{SP}}{K_{a2}} \times [H^+] = 0, \quad (16)$$

$$\left(\frac{K_{a2}}{[H^+]} + 1 + \frac{[H^+]}{K_{a1}}\right) \times [HCO_3^-]^2 - \Sigma [CO_X]_{W0} \times [HCO_3^-] - \frac{K_{SP}}{K_{a2}} \times [H^+] = 0.$$
(17)

Now, we have two quadratic equations with two unknown concentrations ($[HCO_3^-]$ and $[H^+]$). The concentration of $[HCO_3^-]$ is a function of the concentration of protons $[HCO_3^-] = f([H^+])$, containing only one unknown variable, and the concentration of $[H^+]$ must satisfy the conditions of the charge (8) and mass balance (9) equations. The equilibrium concentrations of the $[HCO_3^-]$ and $[H^+]$ ions are calculated iteratively using various values of the $[H^+]$ concentration [24]. The approach for the qualitative and quantitative description of the equilibrium system $H_2O-CO_2-CaCO_3-NaH_2PO_4$ is based on the systematic method for solving multiple equilibrium discrepancies as presented in Skoog et al. [22].

The final equilibrium concentrations of $[Ca^{2+}]$, $[H_2PO_4^-]$, $[H_2PO_4^{2-}]$, $\sum [CO_x]_W$, and pH are calculated based on the equilibrium concentrations $[HCO_3^-]$ and $[H^+]$. The concentration of H^+ ions taken up by CO_3^{2-} ions $(\Delta [H^+]_{CO_3^{2-}})$ and the concentration of H^+ ions released comes from the dissociation of water $(\Delta [H^+]_{H_2O})$, $H_2CO_3 (\Delta [H^+]_{H_2CO_3})$, and $(\Delta [H_2PO_4^-] (\Delta [H^+]_{H_2PO_4^-}))$.

The carbonate equilibrium system $H_2O-[CO_2]_{W0}$ -CaCO₃ was established by adding 100 mg solid CaCO₃ to 1 L MilliQ water with a known initial concentration of dissolved $[CO_2]_{W0}$. Excess CaCO₃ was added so that some solid CaCO₃ remained undissolved in the system at equilibrium. After stabilizing the pH, NaH₂PO₄ was added in concentrations corresponding to 5 mg·L⁻¹, 10 mg·L⁻¹, 25 mg·L⁻¹, 50 mg·L⁻¹, and 100 mg·L⁻¹ of phosphorus, respectively, with molar concentrations of phosphorus ranging from 0.161 to 3.23 mmol·L⁻¹.

Initially, water was purged with pure N_2 gas to reduce the $[CO_2]_{W0}$ concentration to practically zero. pH values were measured using a Jenway type 3520 pH meter and analysed with Dataway

software. A one-litre closed reaction cell was maintained at 25.0 ± 0.1 °C using a water bath thermostat equipped with pH and temperature sensors (Mettler Toledo InLab[®] Routine and Jenway model 027500 sensors, respectively). The experiments were conducted with 1000 mL of water containing 100 mg CaCO₃ (CAS No. 471-34-1, 99.95% Sigma-Aldrich) and NaH₂PO₄ (CAS No. 7558-80-7, 99.5% Acros Organics, US) concentrations ranging from 5–100 mmol·L⁻¹.

Salts were added to the reaction cell after purging with a specific concentration of $[CO_2]$ using gas bags. At the end of the experiment, samples were filtered through a 0.45 µm membrane filter. $[Ca^{2+}]$ ion concentrations were determined by titration using a 1 mM EDTA solution, while the concentrations of dissolved phosphate species were measured using a Hach-Lange DR2800 spectrophotometer and LCK388 cuvettes.

3. Results and discussion

Observations of pH changes in the studied system revealed a 2–4 unit increase in pH_{W0} when CaCO₃ was added to the equilibrium system. This was observed across a concentration range of $[CO_2]_{W0}$ from approximately 3.39×10^{-5} mmol·L⁻¹ to 33.9 mmol·L⁻¹, as shown in Table 2.

Within the given range of $[NaH_2PO_4]$, adding this salt to the initial closed system $H_2O-[CO_2]_{W0}$ -CaCO₃ ($[CO_2]_{W0} \cong 1.32 \times 10^{-2} \text{ mmol} \cdot \text{L}^{-1}$) resulted in a reduction in pH in the final equilibrium system of $H_2O-[CO_2]_{W0}$ -CaCO₃-NaH₂PO₄, as presented in Table 2. The concentration of protons released from the dissociation of H_2CO_3 decreases when the initial concentration of $[CO_2]_{W0}$ is lower. The majority of H⁺ ions, which will react with CO_3^{2-} , are derived from the dissociation of $[H_2PO_4]$.

The difference between $\Delta[H^+]_{H_2PO_4^-}$ and $\Delta[H^+]_{CO_3^{2-}}$ corresponds to the negative proton transfer value of $\Delta[H^+]_{H_2CO_3}$, which results from the reversible dehydration of H_2CO_3 occurring when $[CO_2]_{W0} \ge 3.39 \times 10^{-7} \text{ mmol} \cdot \text{L}^{-1}$. In this scenario, the equilibrium in the system is governed by the dissociation-dehydration processes of H_2CO_3 and the dissolution of CaCO₃.

The concentration of H⁺ ions released from the dissociation of $H_2PO_4^-$ ions ($\Delta[H^+]_{H_2PO_4^-}$) is equal to $[HPO_4^-]$. In the closed system under investigation, when $[CO_2]_{W0} \ge 2.45 \text{ mmol}\cdot\text{L}^{-1}$, the dissociation of H_2CO_3 predominates and produces more protons than the dissociation of $H_2PO_4^-$, causing the pH to decrease below seven. After the addition of NaH_2PO_4 , the range where $\Delta[H^+]_{H_2CO_3}$ has negative values extends to higher values of $[CO_2]_{W0}$. When $[CO_2]_{W0} \cong 2.15 \text{ mmol}\cdot\text{L}^{-1}$ and $c[NaH_2PO_4] \cong 3.23 \text{ mmol}\cdot\text{L}^{-1}$, the corresponding pH value is seven. When $[CO_2]_{W0} \ge 2.45 \text{ mmol}\cdot\text{L}^{-1}$, the values of $\Delta[H^+]_{H_2O}$ become negative because the dissociation of water is suppressed.

The CO_3^{2-} ions released during the dissolution of $CaCO_3$ will accept protons $(\Delta[H^+]_{CO_3^{2-}})$ from two sources: either the reversible dissociation of water $(\Delta[H^+]_{H_2O})$ or H_2CO_3 $(\Delta[H^+]_{H_2CO_3})$. A higher dose of $[CO_2]_W$ increases the solubility of Ca^{2+} ions, lowers pH, and reduces the ratio of $[H_2PO_4^-]_W$ to $[HPO_4^-]$, illustrating the interdependence of carbon and phosphorus equilibria.

Table 2. Equilibrium parameters after adding NaH₂PO₄ (0.161, 0.323, 0.807, 1.61, 3.23 mmol P L⁻¹) to the closed initial system of H₂O-(CO₂)_{w0}-CaCO₃, containing $[CO_2]_{w0} \cong 1.32 \times 10^{-2}$ mmol·L⁻¹ at T = 25 °C

Parameter	Water purged with 1.32×10^{-2} mmol·L ⁻¹ CO ₂ *	[Phosphate-P] with CaCO ₃ 100 mg (1 mmol·L ⁻¹)					
(ions and molecules in mmol·L ⁻¹)		0	0.161	0.323	0.807	1.61	3.23
[Ca ²⁺]	0	0.1044	0.2029	0.3405	0.7162	1.1577	1.7214
[CO ₃ ^{2–}]	0.0025	0.0268	0.0138	0.0092	0.0039	0.0024	0.0016
$[\mathrm{CO}_2]_W[\mathrm{CO}_3^-]$	0.0132	0.0932	0.2045	0.3465	0.7146	1.1180	1.5783
$\sum [CO_X]_W$	0.0157	0	0.0003	0.0015	0.0134	0.0529	0.1567
$[\mathrm{H_2PO_4^-}]$	0	0.1201	0.2186	0.3563	0.7319	1.1734	1.7371
$[H_2PO_4^{2-}]$	0	0	0.0018	0.0100	0.0982	0.4179	1.3689
pH	5.61	0	0.159	0.3130	0.7087	1.1922	1.8611
		9.79	9.16	8.70	8.07	7.66	7.34

* tests without CaCO₃ and NaH₂PO₄

$[PO_4^{3-}]$ mmol·L ⁻¹	$\Delta [H^+]_{H_2O}$	$\Delta[H^+]_{H_2CO_3}$	$\Delta [\mathrm{H^{+}}]_{\mathrm{CO}_{3}^{2-}}$	$\Delta [\text{H}^+]_{\text{H}_2\text{PO}_4^-}$	Theoretical pH	Actual pH
	$mmol \cdot L^{-1}$	mmol·L ⁻¹	$mmol \cdot L^{-1}$	$mmol \cdot L^{-1}$		
0	6.18×10^{-2}	1.59×10^{-2}	-7.75×10^{-2}	0	9.79	9.79
0.161	1.45×10^{-2}	1.54×10^{-2}	-1.89×10^{-1}	1.59×10^{-1}	9.16	9.18
0.323	5.10×10^{-3}	1.42×10^{-2}	-3.32×10^{-1}	3.13×10^{-1}	8.70	8.74
0.807	1.17×10^{-3}	2.31×10^{-2}	-7.12×10^{-1}	$7.09 imes 10^{-1}$	8.07	8.08
1.61	4.43×10^{-4}	3.73×10^{-2}	-1.16	1.61	7.66	7.64
3.23	1.76×10^{-4}	1.41×10^{-1}	-1.72	1.86	7.34	7.22

Table 3. Proton balance in the H₂O-[CO₂]_w-CaCO₃-NaH₂PO₄ system for various dissolved phosphorus concentrations



Fig. 2. Experimentally measured pH values of the closed equilibrium system $H_2O-[CO_2]_W-CaCO_3-NaH_2PO_4$ at a concentration range of $[CO_2]_{WO} \cong 3.39 \times 10^{-5}$ to 3.39×10^1 mmol·L⁻¹ at T = 25 °C.

Consequently, protons bind to CO_3^{2-} ions, increasing the concentration of OH⁻ ions and raising pH to a theoretical maximum of 9.79. The experimentally measured maximum pH of 9.79 ± 0.03 occurred at $[CO_2]_{W0} \cong 3.39 \times 10^{-5}$ mmol·L⁻¹, aligning closely with the theoretically calculated value.

After stabilizing the pH of the equilibrium system $H_2O-[CO_2]_{W0}$, following the addition of CaCO₃, we added NaH₂PO₄ at concentrations of 5 mg P L⁻¹, 10 mg P L⁻¹, 25 mg P L⁻¹, 50 mg P L⁻¹, and 100 mg P L⁻¹ to 1 L of liquid phase, which approximates the concentration range of phosphate phosphorus typically found in reject water [25–27].

In Fig. 2, at an initial concentration of $[CO_2]_{W0} \cong 3.39 \times 10^{-5} \text{ mmol} \cdot \text{L}^{-1}$, the pH decreased from 9.79 ± 0.03 to 7.22 ± 0.02 upon adding NaH₂PO₄ to the system H₂O–(CO₂)_{W0}–CaCO₃, indicating a pH decline of 2.57 ± 0.02 units.

Table 3 shows that the experimentally measured final stabilized pH values of the closed equilibrium system H_2O – $(CO_2)_W$ –CaCO₃–NaH₂PO₄ closely approximated the theoretical values, with slight differences possibly attributable to the accuracy of the gas flow control system. This methodology can be extended to more intricate systems involving multiple acid-base equilibria, such as those with phosphoric acid. Thus, the reliability of our developed theoretical model was confirmed through the outcomes of this experimental study.

4. Conclusions

In this study of modelling the closed system $H_2O-(CO_2)_W$ -CaCO₃-NaH₂PO₄, we based our analysis on the system $H_2O-(CO_2)_W$ -CaCO₃ and enhanced it using principles of proton transfer. The developed model calculated proton transfer parameters, pH, and equilibrium concentrations of ions and molecules in the liquid phase.

In forming the equilibrium system $H_2O-(CO_2)_W$ –CaCO₃ through the dissolution of CaCO₃, CO_3^{2-} ions release and accept protons (Δ [H⁺]_{CO₃}²⁻) from two sources: the dissociation of water (Δ [H⁺]_{H₂O₃) or H_2CO_3 (Δ [H⁺]_{H₂CO₃}). A lower initial concentration of [CO₂]_{W0} results in fewer protons produced through the dissociation of H_2CO_3 . Therefore, water dissociation predominates at the studied}

 $[CO_2]_{W0}$ concentrations, contributing the majority of protons to the equilibrium system. Consequently, protons bind to CO_3^{2-} ions, leading to an increase in the pH of the formed equilibrium system.

In the system H₂O-[CO₂]_w-CaCO₃-NaH₂PO₄, at studied initial concentrations of [CO₂]_{w0}, H₂PO₄⁻ ions served as the primary source of protons. The experimentally measured data closely matched the theoretically calculated values.

An equilibrium model is applicable when reactions occur rapidly or have sufficient time to achieve equilibrium. The developed model for the closed equilibrium system $H_2O-(CO_2)_{WO}$ CaCO₃-NaH₂PO₄ enables determination of component concentrations in the water phase across a wide range of natural and industrial conditions, utilizing the known concentration of $[CO_2]_{W0}$. This model facilitates the assessment of anthropogenic impacts on the environment and can be utilized by environmental technologists in water and wastewater treatment. The model is instrumental in advancing innovative methods for measuring $[CO_2]_W$ and $[HPO_4^{2-}]$ concentrations in water, supporting the development of new environmental technologies.

The theoretical and practical aspects of this study are fundamentally congruent, indicating the successful development of a functional model. Given the significant interest and demand for modelling ecological, chemical, and biological systems, there is ample opportunity to explore and construct more intricate models that can enhance wastewater treatment methodologies.

Data Availability Statement

All research data are contained within the article and can be shared upon request from the authors.

Acknowledgements

The study was supported by the following projects, for which we gratefully acknowledge the fundings: SARASWATI 2 'Identifying Best Available Technologies for Decentralized Wastewater Treatment and Resource Recovery for India', SLTKT20427 'Sewage Sludge Treatment from Heavy Metals, Emerging Pollutants and Recovery of Metals by Fungi', and the University of Tartu Development Fund PLTKT ARENG53 'Isolation and Cultivation of Fungal Strains (including Molds) in Laboratory Conditions for Heavy Metal Recovery from Sewage Sludge'. Additional support was provided by the projects 'Improving Quality of BSR Waters by Advanced Treatment Processes', CELSA AnaCat 'In-situ Catalytic Bioconversion of Pharmaceutically Active Compounds in Wastewater', NANAQUA, NEUTEN, PLANTED Enlight, 'Greenhouse Gas Mitigation by Anaerobic Ammonium Oxidation' by EIC funds, ARENG5, ARENG73 projects and COST Actions CA20101, CA20127, CA20138, CA22102, CA22110, CA22123, CA22162, CA21146, and CA20133. The publication costs of this article were covered by the University of Tartu and the Estonian Academy of Sciences.

References

- 1. Tenno, T., Uiga, K., Mashirin, A., Zekker, I. and Rikmann, E. Modeling closed equilibrium systems of H2O-dissolved CO2solid CaCO₃. J. Phys. Chem. A., 2017, 121(16), 3094-3100.
- 2. Tenno, T., Rikmann, E., Zekker, I., Tenno, T., Daija, L. and Mashirin, A. Modelling equilibrium distribution of carbonaceous ions and molecules in a heterogeneous system of CaCO3water-gas. Proc. Estonian Acad. Sci., 2016, 65(1), 68-77.
- 3. Tenno, T., Rikmann, E., Uiga, K., Zekker, I., Mashirin, A. and Tenno, T. A novel proton transfer model of closed equilibrium systems of H₂O-CO₂-CaCO₃-NH_X. Proc. Estonian Acad. Sci., 2018, **67**(3), 260–270. https://doi.org/10.3176/proc.2018.3.04
- 5. Tenno, T., Rikmann, E., Zekker, I. and Tenno, T. Modelling the solubility of sparingly soluble compounds depending on their particles size. Proc. Estonian Acad. Sci., 2018, 67(3), 300-302. https://doi.org/10.3176/proc.2018.3.10
- 6. Uiga, K., Tenno, T., Zekker, I. and Tenno, T. Dissolution modeling and potentiometric measurements of the SrS-H2O-gas system at normal pressure and temperature at salt concentrations of 0.125-2.924 mM. J. Sulfur Chem., 2011, 32(2), 137-149. https://doi.org/10.1080/17415993.2011.551937
- 7. Zekker, I., Tenno, T., Selberg, A. and Uiga, K. Dissolution modeling and experimental measurement of CaS-H₂O binary

system. Chinese J. Chem., 2011, 29(11), 2327-2336. https:// doi.org/10.1002/cjoc.201180399

- 8. Uiga, K., Rikmann, E., Zekker, I., Mashirin, A. and Tenno, T. Modelling and experimental measurement of the closed equilibrium system of H₂O-SrS. Proc. Estonian Acad. Sci., 2020, 69(4), 287-297. https://doi.org/10.3176/proc.2020.4.02
- 9. Uiga, K., Rikmann, E., Zekker, I. and Tenno, T. Detection and dissolution of sparingly soluble CaS and SrS particles in aqueous media depending on their size distribution. Proc. Estonian Acad. Sci., 2020, 69(4), 323-330. https://doi.org/10.3176/proc.2020.4.07
- 10. Mandel, A., Zekker, I., Jaagura, M. and Tenno, T. Enhancement of anoxic phosphorus uptake of denitrifying phosphorus removal process by biomass adaption. Int. J. Environ. Sci. Technol., 2019, 16(10), 5965–5978. https://doi.org/10.1007/s13762-018-02194-2
- 11. Zekker, I., Rikmann, E., Tenno, T., Kroon, K., Seiman, A., Loorits, L. et al. Start-up of low-temperature anammox in UASB from mesophilic yeast factory anaerobic tank inoculum. Environ. Technol., 2014, 36(2), 214-225. https://doi.org/10.10 80/09593 330.2014.941946
- 12. Christensen, T. H., Kjeldsen, P., Bjerg, P. L., Jensen, D. L., Christensen, J. B., Baun, A. et al. Biogeochemistry of landfill leachate plumes. Appl. Geochem., 2001, 16(7-8), 659-718. https://doi.org/10.1016/S0883-2927(00)00082-2

21

- Tatsi, A. A. and Zouboulis, A. I. A field investigation of the quantity and quality of leachate from a municipal solid waste landfill in a Mediterranean climate (Thessaloniki, Greece). *Adv. Environ. Res.*, 2002, 6(3), 207–219.
- Kjeldsen, P., Barlaz, M. A., Rooker, A. P., Baun, A., Ledin, A. and Christensen, T. H. Present and long-term composition of MSW landfill leachate: a review. *Crit. Rev. Environ. Sci. Tech.*, 2002, 32(4), 297–336.
- Rajagopal, R., Massé, D. I. and Singh, G. A critical review on inhibition of anaerobic digestion process by excess ammonia. *Bioresour. Technol.*, 2013, 143, 632–641.
- Zekker, I., Rikmann, E., Mandel, A., Kroon, K., Seiman, A., Mihkelson, J. et al. Step-wise temperature decreasing cultivates a biofilm with high nitrogen removal rates at 9°C in short-term anammox biofilm tests. *Environ. Technol.*, 2016, **37**(15), 1933– 1946.
- Berends, D. H. J. G., Salem, S., van der Roest, H. F. and van Loosdrecht, M. C. M. Boosting nitrification with the BABE technology. *Water Sci. Technol.*, 2005, **52**(4), 63–70.
- Stillinger, F. H. Proton Transfer Reactions and Kinetics in Water. In *Theoretical Chemistry: Advances and Perspectives*, 3. Academic Press, New York, San Francisco, London, 1978.
- 19. Zumdahl, S. S. *Chemical Principles*. 2nd ed. D.C. Heath, Lexington, MA, 1995.
- 20. Chang, R. *Physical Chemistry with Applications to Biological Systems.* 2nd ed. Macmillan Publishing, New York, 1981.

- Pocker, Y. and Bjorkquist, D. W. Stopped-flow studies of carbon dioxide hydration and bicarbonate dehydration in water and water-d2. Acid-base and metal ion catalysis. *J. Am. Chem. Soc.*, 1977, **99**(20), 6537–6543.
- 22. Segal, B. G. *Chemistry, Experiment and Theory*. 2nd ed. John Wiley & Sons, New York, 1989, 363–365.
- Skoog, D. A., West, D. M. and Holler, F. J. *Fundamentals of Analytical Chemistry*. 6th ed. Saunders College Publishing, Fort Worth, Tex, 1992.
- 24. Dean, J. A. *Lange's Handbook of Chemistry*. McGraw-Hill, New York, 1992.
- 25. Brown, T. L. and LeMay, H. E., Jr. *Chemistry: the Central Science*. Prentice Hall, Englewood Cliffs, New Jersey, 1988.
- Zekker, I., Rikmann, E., Kroon, K., Mandel, A., Mihkelson, J., Tenno, T. et al. Ameliorating nitrite inhibition in a lowtemperature nitritation–anammox MBBR using bacterial intermediate nitric oxide. *Int. J. Environ. Sci. Technol.*, 2017, 14, 2343–2356. https://doi.org/10.1007/s13762-017-1321-3
- Zekker, I., Rikmann, E., Oja, J., Anslan, S., Borzyszkowska, A. F., Zielińska-Jurek, A. et al. The selective salinity and hydrazine parameters for the start-up of non-anammox-specific biomass SBR. *Int. J. Environ. Sci. Technol.*, 2023, 20, 12597–12610.
- Zekker, I., Rikmann, E., Tenno, T., Loorits, L., Kroon, K., Fritze, H. et al. Nitric oxide for anammox recovery in a nitrite-inhibited deammonification system. *Environ. Technol.*, 2015, 36(19), 2477–2478.

Prootoni ülekandel põhinev uudne mudel suletud tasakaalulise süsteemi H₂O–(CO₂)_W–CaCO₃–NaH₂PO₄ modelleerimiseks

Ivar Zekker, Oleg Artemchuk, Ergo Rikmann, Kalev Uiga, Laura Daija, Faysal-Al Mamun, Andrejs E. Krauklis, Ruta Ozola-Davidane, Taavo Tenno ja Toomas Tenno†

Uurimistöös käsitletakse uudset prootoni ülekandel põhinevat täiendatud mudelit suletud tasakaalulise süsteemi $H_2O-CO_2-CaCO_3-NaH_2PO_4$ modelleerimiseks matemaatilise iteratsiooni abil, mille aluseks on võetud varasemates uurimustes analüüsitud tasakaaluline süsteem $H_2O-(CO_2)_W-CaCO_3$. Loodud mudel võimaldab arvutada vees teadaolevate CO_2 ja NaH_2PO_4 kontsentratsioonide põhjal kõigi süsteemis moodustuvate ioonide ja molekulide sisalduse ning määrata vastava pH väärtuse. Väljatöötatud mudeli abil saab hinnata inimtegevuse mõju looduslikele veekogudele, optimeerida tööstuslikke protsesse ja modelleerida teisi analoogseid süsteeme.