Oil Shale. 1999. Vol. 16, No. 4 pp. 303-313

## https://doi.org/10.3176/oil.1999.4.03

# REACTIVITY OF OIL SHALE ASHES TOWARDS SULFUR DIOXIDE 3. RECURRENT USE OF ASH FOR FLUE GAS PURIFICATION

# R. KUUSIK, T. KALJUVEE, H. VESKIMÄE

Tallinn Technical University, Department of Basic and Applied Chemistry 5 Ehitajate St., Tallinn 19086 Estonia

# Yu. ROUNDYGIN

St. Petersburg State Technical University 29 Politechnicheskaya St., St. Petersburg 195251 Russia

# A. KELTMAN

*Eesti Energia* SC 1 Estonia Blvd., Tallinn 10143 Estonia

> A comparison of the reactivity of solid wastes formed during electricity production by firing pulverized Estonian oil shale was prepared and some specific characteristics of oil shale ash were pointed out. Some problems related to the selection of the sorbent injection point are discussed. Industrial-scale experiments for desulfurization of flue gases by recurrent use of cyclone ash and mud from an ash hydrotransport water sedimentation pond were carried out, with positive results, at the Baltic Power Plant in co-operation with Eesti Energia SC, St. Petersburg State Technical University and Tallinn Technical University.

### **Reactivity of Different Types of Oil Shale Ashes**

Estonian and Baltic Power Plants (PP) operate on combustion technology based on pulverized firing of Estonian oil shale at atmospheric pressure and at temperatures 1200-1400 °C. Oil shale is characterized by a high content of mineral matter which at the combustion leads to the formation of huge quantities of waste ashes – 5-6 million tons at the production level of the year 1996 [1].

amp
5
of
5
0
ti
• 1000
I
t
ac
1
3
-
0
=
•=
<b>CO</b>
2

es

Parameter	Samples		Ŧ					
the Ba bes the Ba bes the fi the fi	Furnace ash	Superheater ash	Cyclone ash	Electrostatic precipitator ash	Low-temperature furnace ash	Wet ash	Mud	Natural limestone from the Karinu deposit
Content (wt.%): CaO <sub>total</sub>	54.9	52.5	49.1	36.5	45.0	42.0	52.6	52.6
CaO <sub>free</sub> : For the initial samples	20.8	14.8	13.1	10.6	23.8	5.5	1.1	A
For the activated (grinding) samples	22.7	23.6	18.7		30.0		1	
MgO	5.7	5.4	5.1		9.4	1	0.3	1.5
SO <sup>2-</sup>	4.5	7.1	6.9	8.8	18.4	0.0	3.9	0
$d_{1,0}$ , + Fe,0	10.3	11.2	12.6	10.5	6.5	7.8	1.3	0.3
SiO2	21.1	22.3	23.5	36.5	12.9	19.8	1.2	
BET surface area (m <sup>2</sup> /g):	0.76	0.83	0.62	0.45	1.91	7.40	8.64	2.98
For the activated samples	4.84	4.02	3.33	.18	3.87	AJ.	1	19( 191. U
Content (wt.%) of the fractional class -4	45 µm:	42.1	58.5	98.3	3.1	28.1	99.1	99.5
For the activated samples	89.3	90.8	89.8	123 14	92.3			1



*Fig. 1.* Scheme of the boiler installation at Baltic PP; points of ash removal: 1 - burning chamber-furnace, 2 –superheater, 3 –cyclone, 4 – electrostatic precipitator; A and B – ash injection points on the level of 11 and 20 m from the bottom of the furnace, respectively

Different types of solid wastes are formed at these power plants. The ash formed during pulverized firing of oil shale (high-temperature ash) is removed in different technological installations – furnace (furnace ash, 10-15 % of the total amount of the ash formed), superheater (superheater ash, 15-25 %), cyclone (cyclone ash, 35-50 %) and electrostatic precipitator (electrostatic precipitator ash, 15-20 %) (Fig. 1). Furnace and superheater ashes are removed by the hydroseparation method, cyclone and electrostatic precipitator ashes – by dry method (dry ash).

Compositions Fig. 2. of the Estonian oil shale ash, the fly ash fractions and their phases: 1 - fly ash fractions selected by density and particle size, 2 - furnace ash, 3 cyclone ash, 4 - electrostatic precipitator ash, 5 - coarse fraction (median diameter 50-120 µm), 6 fine fraction (12-25 µm), 7 - finest fraction (6-9 µm), 8 - glass constituent of the coarse fraction, 9 - glass constituent of the fine fraction, 10 - glass constituent of the finest fraction, 11 - pozzolanic constituent of all ash fractions



All types of ashes are sent by hydrotransport into an open-air deposit (ash field), as a result of which they are hydrated (wet ash). Finally, in the transport-water sedimentation ponds, partial recarbonization of alkali water and precipitation of mud takes place. The mud contains a noticeable amount of secondary calcium and magnesium carbonates.

The ashes removed in different installations differ from each other in chemical composition and physical properties (Table). The data about the chemical and phase composition of the ashes has been generalized by V. Kikas [2] and presented in Fig. 2. The ash from the electrostatic precipitator, compared to the other types of ash, has the lowest content of free CaO and the highest content of alite and belite, and has found extensive use in cement industry [2]. Cyclone ash is well known as a lime-containing material for neutralizing acid soils. Wet ash and mud contain hydrated CaO and increased amounts of carbonates, 5.7 and 35.2 %, respectively.

The chemical activity of oil shale ashes causes serious fouling and corrosion problems for the heat transfer surfaces of the boiler [3], especially in the pulverized firing process, but it also enables the binding of about 80 % of sulfur dioxide into the solid phase (ash) formed at combustion [4].

At low-temperature combustion of oil shale (at 850-900 °C), when the circulating fluidized bed combustion (CFBC) technology is used, the corrosion problems are not so evident [5] and the SO<sub>2</sub> binding is almost completed [6].



Fig. 3. Scheme of the dry enrichment of oil shale ashes

Due to the composition of oil shale ash, particularly to their free CaO content (up to 23 %), they could be potential sorbents for deeper removal of SO<sub>2</sub> from flue gases [7, 8]. High reactivity towards sulfur dioxide [8, 9] and the possibility of using oil shale ash as a sorbent for binding SO<sub>2</sub> from the flue gases by dry method have been pointed out [9-11]. In addition, our investigations have shown that dry ashes can be easily activated [12, 13] and enriched. It was possible, by means of grinding and air separation, on the basis of furnace ash to produce a sorbent with particle size -100  $\mu$ m, in which the content of free CaO was 26.6 % and the yield of free CaO - 82.9 % (Fig. 3).



*Fig.* 4. SO<sub>2</sub>-binding capacity of ashes at 700 °C after 2-min contact. Legend: 1 -furnace ash; 2 - superheater ash; 3 - cyclone ash; 4 - electrostatic precipitator ash; 5 - low-temperature furnace ash; 6 - wet ash; 7 - mud; 8 - natural limestone from the Karinu deposit; G - grinding; G + H - grinding and subsequent hydration

Figure 4 presents the results of investigations aimed at estimation of the reactivity of different waste materials formed at an oil shale firing power plant, including the activated forms of ashes and, for comparison, ashes formed by using CFBC technology as well as natural limestone from the Karinu deposit. The SO<sub>2</sub>-binding capacity (mg SO<sub>2</sub> per 100 mg sample or mg SO<sub>2</sub> per 100 mg free CaO) was used as a parameter for comparison [12].

At 700 °C and 2-min contact between the solid and gaseous phases, the initial samples of high-temperature ash bound 10-11 mg SO<sub>2</sub> per 100 mg sample and the initial samples of low-temperature furnace ash and wet ash -17.3 and 21.5 mg SO<sub>2</sub> per 100 mg sample, respectively, or 1.6 and 2.1 times more than high-temperature ashes. Previous activation of low-temperature furnace ash by grinding, as well as activation of high-temperature ash by grinding and subsequent hydration, increased the binding capacity of the ashes approximately two times – up to 32 and 23 mg SO<sub>2</sub> per 100 mg sample, respectively. The binding capacity of mud (37 mg SO<sub>2</sub> per 100 mg sample) is comparable to that of natural limestone (41 mg SO<sub>2</sub> per 100 mg sample) (Fig. 4).

Thus, the laboratory investigations proved the following:

• The properties and reactivity of dry ashes depend on the place of technological installation for their removal

- Dry ashes can be activated towards SO<sub>2</sub> and the content of free CaO can be increased by enrichment
- Wet ashes are already activated, carbonaceous muds are highly reactive towards SO<sub>2</sub>
- By their reactivity, the ashes formed in CFBC conditions are similar to wet ashes

The results of the experiments confirmed the perspective of using some types of oil shale ashes and mud in recycle to reduce  $SO_2$  pollution during combustion of powdered oil shale.

### **Selection of the Sorbent Injection Point**

The efficiency of the dry purification method of flue gases depends on the following parameters:

- Type and amount (ratio of CaO/S) of the sorbents injected
- Mechanical, physical and chemical properties of the sorbent and its reactivity
- Temperature and duration of the sorbent SO<sub>2</sub> contact

As these factors are almost completely determined by the technological parameters at the ash injection point, we will discuss the named factors more closely, taking into consideration the specific properties of oil shale ashes.

#### The Kind and Amount of Sorbent Used

During coal combustion the sorbent (usually limestone or dolomite) is fed into the burner together with the fuel. The amount of the sorbent is described by the mole ratio CaO/S (CaO/SO<sub>2</sub>) in the mixture. The optimum value of the index is 2-3. This value is calculated considering the total CaO content in the sorbent.

By using oil shale as a fuel, it must be considered that this value depends on the content and composition of the mineral part of oil shale ranging from 8 to 10. In such case, injection of moderate amounts of the sorbent with the fuel into the burner is not effective, injection of a large amount of the sorbent, however, is too expensive. This is why the sorbent has to be injected into the boiler on its upper level, where the particles of the oxides formed during decomposition of the carbonaceous part of oil shale are already covered with sulfate crust and the SO<sub>2</sub> binding is practically completed. The SO<sub>2</sub> concentration being lower there (~2000-3000 mg per nm<sup>3</sup>), compared to its content on the lower level of the boiler, a value of the mole ratio CaO/S = = 2-3 could be achieved by using a relatively small amount of the sorbent.

When injecting oil shale ash sorbents, the question is which kind of content – that of total CaO or of free CaO – is more characteristic for the calculation of this value. The total CaO content in oil shale ash is on the level of 49-58 %, the content of free CaO depends on the type of ash and is

13-20 % (abs.). As the bound CaO (e.g. calcium silicates) takes part in the binding of SO<sub>2</sub> [9, 14, 15], some authors take the total content of CaO as a basis. In our opinion, oil shale sorbents are better characterized by the ratio of free CaO/S. The point is that, based on the kinetic parameters, during a short-time contact between the sorbent and SO<sub>2</sub>, in the first order, free CaO as the more active compound takes part in the reaction with SO<sub>2</sub>.

### **Specific Properties and Reactivity of Sorbent**

In the dry purification process of flue gases the quality of the limecontaining sorbent is determined by the content of the basic compounds (CaO + MgO) and by the content of the impurities as well as by physical properties of the samples (particle size, porosity and distribution of pores by their diameter). The indicator characterizing the last parameters is the specific surface area of the samples.

When oil shale ash sorbents are used following circumstances must be taken into cosideration:

 Firstly, it is important to increase the specific surface area by activating the samples (Table). The dependence of SO<sub>2</sub>-binding rate on the specific surface area of oil shale ashes is shown in Fig. 5. The hydration of the ashes favours the increase of their binding capacity by the formation of a new phase – Ca(OH)<sub>2</sub>. The CaO, originated from Ca(OH)<sub>2</sub> is more active towards SO<sub>2</sub> than that from CaCO<sub>3</sub> [16].



- *Fig.* 5. Dependence of  $SO_2$ binding rate W of different ashes on the specific surface area (SSA) of the sorbents (700 °C, 15 sec)
- Secondly, sintering of ashes causing a decrease in the specific surface area and, therefore a decrease in the binding capacity of oil shale ash, is noticeable at 800-850 °C (Fig. 6). Thus, the temperature in the contact zone between the solid and gaseous phases should not exceed this temperature too much.
- Thirdly, oil shale ash sorbents differ from limestone in containing sulfates (mainly CaSO<sub>4</sub>) already before injection into the boiler or the gas tract. The decomposition temperature of CaSO<sub>4</sub> and re-emission of SO<sub>2</sub> depends on the composition and quantity of impurities and on the

composition of gas phase. Due to the content of iron, aluminium and silicon compounds, the thermal decomposition of sulfates (re-emission of SO<sub>2</sub>) starts at 950-1000 °C [17, 18], in the presence of reducers (CO) it might start at 600 °C [19, 20].



*Fig. 6.* Dependence of the specific surface area of oil shale ashes on the processing temperature: *1* - circulating ash of CFBC test facility, *2* - cyclone ash of Baltic PP

### Duration and Temperature of the Sorbent - SO<sub>2</sub> Contact

Generally, during heterogeneous interaction of the sorbent and  $SO_2$ , high efficiency of the sorbent utilization could be obtained in the case of a long-time contact. An increase in the temperature increases the  $SO_2$ -binding rate.

At pulverized firing the residence time of the oil shale particles in the burner is approximately ten seconds. It was shown by Roundygin et al. [11] on the basis of the TP-17 boiler that by selecting the optimum point, the conditions for injection of the sorbent and by considering the aerodynamics of the gases inside the combustion chamber, it is possible to multiply the residence time of the particles in the burner. Thus, as the nonoptimum temperature (too high temperature) in the sorbent injection point could eliminate the expected effect, the right points and efficiency of recycling the sorbent must be checked in the industrial-scale experiment only.

#### **Results of Industrial-Scale Experiments**

The industrial-scale experiments were carried out in the boiler TP-17 at the Baltic PP (oil shale consumption 60 t/h, steam production 150 t/h). The results are briefly described in [21].

Considering the acceptable physical properties, the high quantity of free CaO in the cyclone ash fractions with particle size over 100  $\mu$ m and the fact that only about 5 % of the initial CaO may react in the boiler furnace, cyclone ash was used as an initial material for sorbent production in the first experiment. In the second stage of the experiments, dried mud was used.

Cyclone ash, previously ground in a ball mill, was supplied into the burning chamber of the boiler through 6-8 nozzles to ensure a uniform overall filling of the interior furnace space with the sorbent. The flow rate of transporting air and the cross-section of the nozzles ensured a minimum velocity of 20-25 m/s of the sorbent at the entry.

The injection points A and B (Fig. 1) on the level of 11 and 20 m from the bottom of the furnace, in the temperature range of the flue gases of 1000-1100 and 850-900 °C, respectively, were chosen considering the aero-dynamics and temperature profile of the burning chamber, the results of simulation of the residence time of the particles in the burning chamber (by variant A up to 30 seconds) and the SO<sub>2</sub> capture data [11]. The feed of the recycled ash varied in the range of 6-10.5 t/h (21-32 % of the total ash formed, mole ratio of free CaO/S = 1.6-2.8). Sorbent mud was injected into the boiler as a mixture with oil shale (mass ratio 1 : 6).

By injection of cyclone ash on level A there were no positive results – the  $SO_2$  concentration in the flue gases neither decreased nor depended on the amount of the ash injected. Obviously, the sorbent was injected into the circulation zone of gases on a low level of  $SO_2$  concentration. At temperatures 1000-1100 °C the ash was inactivated and lost its  $SO_2$ -binding ability. By using injection level B the concentration of  $SO_2$  in the gas phase decreased, depending on the amount of the recycled ash, by 17-34 % (an increase in the binding efficiency from 57 to 72 %).

Using the mud injected into the boiler as a mixture with oil shale, a decrease down to 40-50 % and less in the SO<sub>2</sub> concentration in the flue gases was achieved (from 2200-2100 to 1300-1000 mg SO<sub>2</sub> per Nm<sup>3</sup>). The total SO<sub>2</sub>-binding efficiency increased from 68 % to 85 %.

## Conclusions

- 1. The solid wastes formed at the Estonian oil shale fired power plants have a considerably high remaining reactivity towards SO<sub>2</sub>, therefore they could be accepted as potential sorbents for a deeper desulfurization of flue gases by dry method. They differ from each other in physical properties, chemical composition and reactivity towards SO<sub>2</sub>. By using different methods of activation (and dry enrichment), it is possible to double the SO<sub>2</sub>-binding capacity of the initial samples.
- 2. It was shown that at recurrent use of oil shale ash it is very important to observe the temperature field in the furnace and the specific characteristics of the ash, in order to avoid their sintering-inactivation and decomposition of the sulfates contained in the recycled ashes. Interaction between the recycled ash and the gaseous phase must take place in the low SO<sub>2</sub> concentration zone to guarantee the high CaO/S ratio by low consumption of the sorbent. The temperature of the gases at the sorbent injection level (point) is not to exceed 900-950 °C.
- 3. The results of the industrial-scale experiments at the Baltic PP confirmed the efficiency of using cyclone ash and mud in recycle to reduce the  $SO_2$ emission by combustion of powdered oil shale for electricity production. An increase in the  $SO_2$ -binding efficiency from 57 % to 72 % was achieved by injecting the previously ground cyclone ash. The feeding of

dried mud into the burner in a mixture with oil shale decreased the  $SO_2$  concentration in the flue gases down to 50 % and less. Thus, in that case the  $SO_2$ -binding efficiency increased from 68 % to 85 %. Considering the nonoptimum temperature level for the feeding point of mud, a higher efficiency of  $SO_2$ -binding could be expected.

4. On the basis of the data collected it is possible to prepare the initial data for designing a flue gas purification technology by using oil shale ash sorbents. The technology for preparing the sorbent by grinding on an industrial scale was checked. The SO<sub>2</sub>-binding technologies by using wet ash or mud need further investigation.

#### Acknowledgements

The authors express their gratitude to *Eesti Energia* SC for the partial funding of the present work, to Dr. Dmitri Yegorov for his assistance in arranging the measurements, and to Mr. Ülo Nagland for his help and an excellent organization of the industrial-scale experiments at the Baltic PP.

### REFERENCES

- Estonian Environment 1996 / Ministry of the Environment of Estonia. Tallinn, 1997.
- Kikas W. Composition and binder properties of Estonian kukersite oil shale ash // Zement-Kalk-Gips. 1997. Vol. 50, No. 2. P. 112-116.
- Fouling and Corrosion in Steam generators / Eds. D. Savic, I. Öpik, B. Kidric. Institute of Nuclear Science, Vinca; Tallinn Technical University. - Beograd, 1980.
- Ots A. Utilization of Estonian oil shale as high carbonates content fuel // 3rd Intern. Conf. on Coal Utilization Science and Technology, Bucharst, May 6-7, 1998. Session 2, Paper 16.
- Arro H., Prikk A., Kasemetsa J. On the fouling of heat transfer surfaces of CFB oil shale boiler // Oil Shale. 1997. Vol. 14, No. 3 Special. P. 218-224.
- Arro H., Prikk A., Kasemetsa J. Circulating fluidized bed technology Test combustion of Estonian oil shale // Oil Shale. 1997. Vol. 14, No. 3 Special. P. 215-217.
- Marier P., Dibbs H. P. The catalytic conversation of SO<sub>2</sub> to SO<sub>3</sub> by fly ash and the capture of SO<sub>2</sub> and SO<sub>3</sub> by CaO and MgO // Thermochimica Acta. 1974. Vol. 78. P. 155-165.
- Kaljuvee T., Kuusik R., Vilbok H. High-temperature adsorption of sulfur dioxide // Kemia-Kemi. 1990. Vol. 17, No. 10B. P. 993-994.
- Kaljuvee T., Kuusik R. Desulfurization of gases by oil shale ash // Oil Shale. 1993. Vol. 10, No. 1. P. 34-44.

- Kuusik R., Kaljuvee T., Trikkel A., Maarend J., Roundygin Yu. Dry desulfurization of flue gases by Estonian lime-containing materials // Proc. 10th World Clean Air Congr. Espoo, Finland, May 28-June 2, 1995. Helsinki. Vol. 1. P. 037.
- Rundygin Yu., Alfimov G., Rundygin A., Grigorjev K., Maarend J., Arkhipov Yu., Kuusik R. Possibilities of deeper desulfurization of flue gases by oil shale ash components in different burning technologies // Oil Shale. 1997. Vol. 14, No. 2. P. 115-131.
- Kaljuvee T., Trikkel A., Kuusik R. Reactivity of oil shale ashes towards sulfur dioxide. 1. Activation of high-temperature ashes // Oil Shale. 1997. Vol. 14, No. 3. P. 393-408.
- Kuusik R., Kaljuvee T., Trikkel A., Arro H. Reactivity of oil shale ashes towards sulfur dioxide.
  Low temperature ashes formed by using CFBC technique // Oil Shale.
  No. 1. P. 51-63.
- 14. Arro H. H., Mahlapuu A. J., Ratnik V. E. The role of bound CaO in the process of forming ash sediments // Proc. Conf. on Processes Occurring in the Fuel Mineral Part. Tallinn, 1969. P. 95-102 [in Russian].
- 15. Külaots I., Ots A., Yrjas P., Hupa M., Beckman P. Sulfation of Estonian and Israeli oil shale ashes under atmospheric and pressurized combustion conditions // Oil Shale. 1997. Vol. 14, No. 3 Special. P. 265-283.
- 16. Bruce K. R., Gullett B. K., Beach L. O. Comparative SO<sub>2</sub> reactivity of CaO derived from CaCo<sub>3</sub> and Ca(OH)<sub>2</sub> // AIChE Journal. 1989. Vol. 35, No 1. P. 37-41.
- Kuusik R., Taremaa Ü. About thermal processing of phosphogypsum from different forms of phosphate raw material // Journal of Applied Chemistry. 1988. No 3. P. 512-517 [in Russian].
- 18. Kuusik R., Veiderma M. Thermal processing of phosphogypsum // Proc. 3rd Intern. Symp. on Phosphogypsum, Orlando, Florida, Dec. 1990. P. 267-279.
- Lingefelt A., Leckner B. SO<sub>2</sub> capture in fluidized bed boilers: re-emission of SO<sub>2</sub> due to reduction of CaSO<sub>4</sub> // Chem. Eng. Sci. 1989. Vol. 44, No. 2. P. 207-213.
- 20. Kuusik R., Saikkonen P., Niinistö L. Thermal decomposition of calcium sulfate in carbon monoxide // J. of Therm. Anal. 1985.Vol. 30. P. 187-193.
- 21. Kuusik R., Kaljuvee T., Trikkel A., Roundygin Yu., Alfimov G., Jegorov D., Maarend J. Flue gas desulfurization at oil shale fired plants // 4th Intern. Symp. And Exib. "Environmental Contamination in Central and Eastern Europe", Warsaw, Poland, Sept. 15-17, 1998. Symp. Proc. Tallahassee, Florida, USA: Florida State University, 1999. Paper No. 272. CD-ROM.

Presented by I. Öpik Received February 16, 1999